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Covalent organic polymer functionalization of activated carbon surfaces through acyl chloride for environmental clean-up

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Abstract

Nanoporous networks of covalent organic polymers (COPs) are successfully grafted on the surfaces of activated carbons, through a series of surface modification techniques, including acyl chloride formation by thionyl chloride. Hybrid composites of activated carbon functionalized with COPs exhibit a core-shell formation of COP material grafted to the outer layers of activated carbon. This general method brings features of both COPs and porous carbons together for target-specific environmental remediation applications, which was corroborated with successful adsorption tests for organic dyes and metals.

Keywords: Activated carbon, Covalent organic polymers, Water treatment, Acyl chloride, Surface grafting

1. Introduction

As water scarcity around the globe continues to deepen and water treatment becomes increasingly more crucial for the health and development of societies on every corner of the planet [1], the necessity for new, effective, and affordable treatment materials also becomes imperative. Cheap and efficient technologies employing novel materials are emerging across a wide range of disciplines, from nanomaterials [2–4], to biomimetic membranes [5,6], to modifications on classic technologies like activated carbon [7–10]. To date, the standard cheap adsorptive material in water treatment remains to be activated carbon. Activated carbon is, and has been for decades, actively employed in the adsorption of many common contaminants, such as pesticides [11], pharmaceuticals and endocrine disruptors [12], heavy metals [13], inorganic ions [14], among others. Although the versatility of activated carbon to adsorb a wide range of compounds is generally advantageous, it also becomes a limitation when there is a need for selectivity in treating a particular contaminant. Consequently, various modifications have been made on activated carbon to achieve specific advantages for numerous particular applications. These modifications have been treatment with acids, bases, external media impregnation, heat, microwaves, ozone, plasma, and bioadsorption [15,16]. Therefore, functionalization of this traditional material with novel compounds can expand its use to an even wider range of pollutants, targeting specific contaminated water sources. One of these novel materials is that termed covalent organic polymers (COPs). The benefit of these polymers is that they are extremely stable in many harsh conditions (e.g. high temperatures and boiling water) [17], and have already been proven efficient for environmental applications, such as gas capture [17–19], solvent uptake [20], and groundwater remediation [21].

However, full-scale applications of these COPs for environmental remediation have yet remained elusive. Efficient purification of large volumes of water or air requires that the contaminated stream can be drawn through a column containing highly functional materials, on the order of millimeters in size; in order to achieve

economical synergy between contaminant/column contact time and energy demand. To date, COPs are nano- to micrometer in scale [17,18,21,22] or in the form of macroscale solidified gel-like materials [20]. Columns built with particles of such a small size are rendered unpractical for typical remediation purposes, as they yield very high back pressures, as is the case with high performance liquid chromatography (HPLC) columns; and the energy to overcome this phenomenon would be too great for large-scale operation. Additionally, polymerizing the solidified gel-like polymers into millimeter-scale particles would still be limited in treatment efficacy due to slow intra-polymeric diffusion and insufficient contaminant contact time with the column material. Thus, supporting COPs on a cheap backbone material provides an ideal solution for creating a cost effective purification process.

Herein, we report the functionalization of COPs onto the surface of granular activated carbon (GAC); through a series of surface modification techniques, followed by the synthesis of a COP “shell” around the carbon granule. Activated carbon, established as one of the cheapest and most effective environmental remediation materials of all time [23], provides the optimal base material for the attachment of COPs. As it is large enough to be able to be used in a low energy demand packed-bed column [24] and has a compatible chemistry for the grafting of polymers to its surface. A first of its kind, activated carbon with a porous polymer functionalized shell provides a robust and re-generable material with the durability and versatility for a wide range of environmental applications. By developing a cheap functionalized activated carbon with a novel porous polymer matrix, it becomes possible to drastically increase the target contaminant range for water treatment, while still remaining affordable.

In this study, we successfully functionalized the surface of GAC with the porous matrix of COP material, specifically COP-19 [25]. Through an array of characterization techniques, it was confirmed that not only was the COP material present, but was in fact grafted to the surface of the carbon materials in a shell-like fashion. Given this successful functionalization (surface polymerization) method, it is now possible to develop a

carbon/COP “library” of various different polymers attached to GAC. This library will aim at targeting specific environmental applications, depending on the polymer; from heavy metal up-take, to organics adsorption, to impregnation with reactive iron for contaminant degradation, to gas-capture and biogas purification operations. Subsequently, initial adsorption studies on an azo dye, cadmium, and iron prove feasibility for the developed material to be utilized in environmental applications.

2. Materials and methods

Acid oxidation of the carbon was performed as follows. 25 g of Samchun carbon (AC) was combined with 250 mL of concentrated nitric acid (HNO_3 , 60%) at 100 °C under reflux for 24 h. The carbon was washed thoroughly with copious amounts of MilliQ deionized water, until the rinsate reached a neutral pH. This material (AC-Ox) was then dried in a vacuum oven for 12 h at 110 °C. Acyl chlorination of AC-Ox was performed as follows. 2.5 g of oxidized carbon (AC-Ox) was mixed with a 2:1 mixture of dichloromethane (CH_2Cl_2), 100 mL, and thionyl chloride (SOCl_2), 50 mL, and refluxed for 24 h in a nitrogen atmosphere. The solvents were removed from the material via rotary evaporation to afford the acyl chloride modified carbon (AC-Thio). Melamine attachment to AC-Thio was performed as follows. Avoiding contact with air, approximately 2.5 g of acyl chloride modified particles (AC-Thio) were immediately treated with a solution of 375 mg of melamine ($\text{C}_3\text{H}_6\text{N}_6$) dissolved in 150 mL of dimethyl sulfoxide ($(\text{CH}_3)_2\text{SO}$) and 2.5 mL of diisopropylethylamine (DIPEA) that had been sonicated until complete dissolution had taken place, and heated to 120 °C for 24 h in a nitrogen atmosphere. After filtration, the particles of melamine attached carbon (AC-Mel) were washed thoroughly three times each with dimethyl sulfoxide, then MilliQ deionized water, and finally ethanol; then dried in a vacuum oven for 12 h at 110 °C. Covalent organic polymer attachment to AC-Mel was performed as follows. 500 mg of melamine and 800 mg of terephthalaldehyde was combined with 150 mL of dimethyl sulfoxide, sonicated until complete dissolution took place, and placed into a flask with a nitrogen atmosphere. 1000 mg of the melamine attached carbon particles (AC-Mel) were mixed into the solution, maintaining the nitrogen atmosphere. This mixture was

stirred under reflux for 48 h. After filtration, the particles of polymer attached carbon (AC-COP) were vigorously washed thoroughly three times each with dimethyl sulfoxide, then acetone, then MilliQ deionized water, and finally ethanol; then dried in a vacuum oven for 12 h at 110 °C. Further information about the chemicals and equipment, and adsorption tests, can be found in the Supplementary material, section S1 and S2.

3. Results and discussion

Initial surface modification methods of the activated carbon were inspired from previously reported literature of surface modification using thionyl chloride targeted at graphene [26] and carbon black [27], as well as an amino-functionalization method for carbon nanotubes [28]. Various types of GAC were tested for the ability to be effectively oxidized at the surface to introduce carboxylic groups, and it was discovered that the lack of impurities in a pristine GAC (i.e. Norit[®] activated carbon) made it difficult to break the bonds of the carbon and subsequently change the surface chemistry enough to allow for polymer grafting. Moreover, when using a GAC with too many impurities (i.e. Sigma activated charcoal), the surface area was too low and particle structure broke down too easily during modification. For mass loss, due to impurities on the carbons, determined by thermogravimetric analysis (TGA) of the different carbons, see Supplementary material, section S3. Consequently, a moderately pristine GAC (i.e. Samchun activated carbon) that still possesses a high surface area was found to be the optimal starting point for the reaction. Carbon was first oxidized with concentrated nitric acid (HNO₃, 60%) at 100 °C for 24 h (AC-Ox). Once highly saturated with carboxylic groups, the carbon was treated with a 2:1 mixture of dichloromethane (CH₂Cl₂) and thionyl chloride (SOCl₂) under reflux for 24 h (AC-Thio), converting the carboxylic groups to the extremely reactive acyl chloride substituent. After drying by rotary evaporation, the AC-Thio was used immediately in the next reaction step, without characterization, in order to avoid hydrolysis of acyl chloride by air or moisture. AC-Thio particles were then treated with a solution of melamine dissolved in dimethyl sulfoxide (DMSO) at 120 °C for 24 h (AC-Mel). At this point, melamine is grafted onto the surface of the acyl chloride containing carbon, by making amide bonds. Using the amine groups

that melamine possesses, attachment of a covalent organic polymer (COP) shell around the GAC is thus possible. The final surface polymerization of attaching a COP matrix is employed by previously reported methods utilizing Schiff base chemistry with melamine and terephthalaldehyde as the core monomers (AC-COP) [22,25]. The entire series of reaction schemes is identified in Figure 1. Due to the vigorous washing strategies after each phase of the grafting procedure, it can be assumed that any unattached monomers or polymers are no longer present (i.e. the high solubility of melamine in water and terephthalaldehyde in ethanol will wash away any residuals), and therefore characterization data reflects only the carbon and grafted materials.

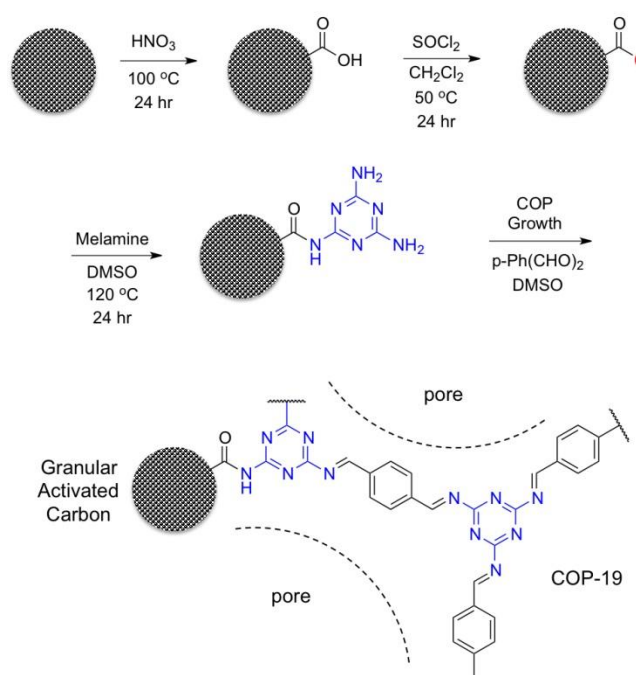


Fig. 1. Step-wise evolution of the COP attachment to the surface of activated carbon (illustrated as large grey object, not to scale).

Fourier transform infrared spectroscopy (FTIR) was carried out at each point in the series of reaction steps producing the carbon-COP composite (Figure 2A). As expected, the raw activated carbon exhibits no distinguishable peak pattern in the spectra, confirming the lack of any significant functional groups. Following

the first reaction step with acid oxidation (AC-Ox), two major peaks appear at 1710 cm^{-1} and 1210 cm^{-1} , which can be attributed to C=O and C-O, respectively, of the stretching of carboxylic groups formed on the carbon surface. Another peak found at 1520 cm^{-1} , correlates to nitro groups formed during the oxidation. After the reaction in the third step with the grafting of melamine (AC-Mel), the appearance of peaks at 1580 and 1000 cm^{-1} correlate to N-H and C-N, respectively. These stretching frequencies verify the existence of secondary amine groups stemming from the acyl chloride on the carbon surface. Ultimately, the final product (AC-COP) following the reaction in the fourth step with the COP attachment, yields a multitude of peaks at 1720 , 1530 , 1460 , 1340 , 1190 , 980 , 860 , and 800 cm^{-1} . These peaks are identical to the peak pattern found in the pure COP-19, confirming that there is now COP material contained within the sample matrix (Figure 2A, green dashed line).

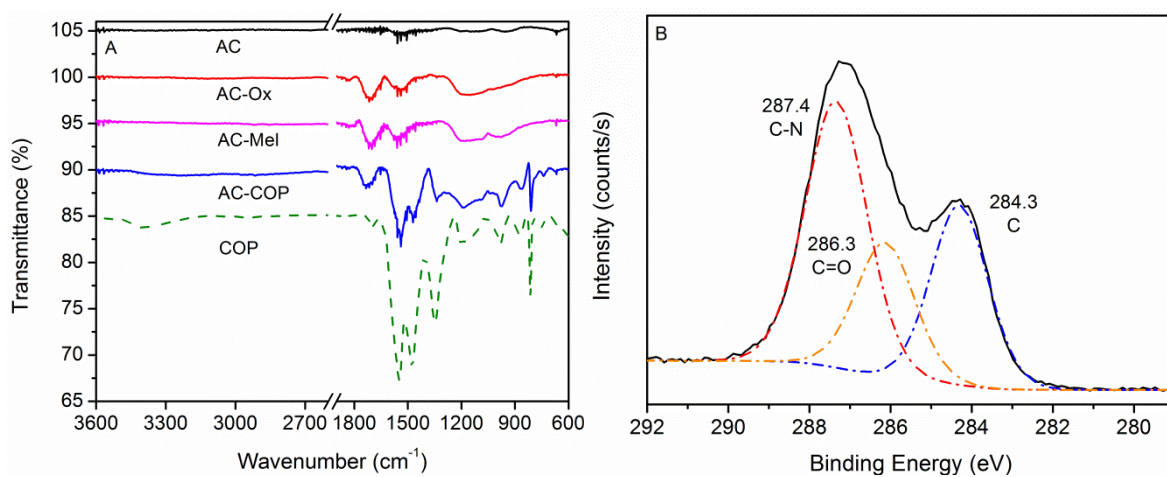


Fig 2. A) FTIR spectra of AC (black), AC-Ox (red), AC-Mel (magenta), AC-COP (blue), and COP-19 (green);
XPS C1s scan of AC-COP.

When elemental analysis (EA) was performed on the carbon particles, there was a clear trend in the quantities of each element present that corresponded very well with the reaction scheme (Table 1A). Initially, as expected, the particles are primarily carbon, with some defects found mostly in the form of oxidation. Following acid

oxidation with nitric acid (AC-Ox), the percentage of oxygen significantly increases; additionally, minor increases in hydrogen and nitrogen takes place. After treatment with thionyl chloride (AC-Thio), and with the attachment of melamine to the carbon surface (AC-Mel), the third step correspondingly sees a substantial increase in nitrogen and hydrogen, due to the added presence of amine groups and aromatic nitrogens; while concurrently seeing a significant decrease in oxygen, as a secondary amine substitutes for the acyl chloride. Furthermore, the COP attachment more than doubles the nitrogen content while lowering the oxygen and hydrogen slightly (AC-COP), matching well with the added presence of a polymer constructed from melamine and terephthaldehyde.

Table 1. Elemental concentrations of carbon particles at each stage of the modification reactions, determined by elemental analysis (A) and X-ray photoelectron spectroscopy (B).

	C	N	O	H	Other
<i>A. Elemental analysis (%)</i>					
AC	82.40	0.58	6.50	1.67	8.85
AC-Ox	55.43	2.16	33.40	2.04	6.97
AC-Mel	61.90	7.57	19.48	4.23	6.82
AC-COP	55.85	17.52	12.43	2.65	11.54
<i>B. X-ray photoelectron spectroscopy (%)</i>					
AC	77.36	0.00	18.58	---	4.06
AC-COP	57.28	32.80	6.36	---	3.56

However, in order to verify that the spectral data from FTIR was indeed indicative of a polymer and its corresponding functional groups on the surface of the carbon, rather than entirely separate co-existing particles, X-ray photoelectron spectroscopy (XPS) was employed, which is limited to a penetration depth of the first 10 nm of a surface. Analyzing the elemental percentages with XPS reveals a clear indication that there is a polymer shell bound to the surface of the carbon, see Table 1B. The best indicator for this is the surface nitrogen, which is non-existent in the original carbon, and is nearly one-third of the total content when the polymer is grafted. Additionally, an in-depth look at the C1s scan of the XPS measurement yields three distinct peaks of C-N, C=O,

and C (Figure 2B), whereas the original carbon contains only the bare carbon peak (see Supplementary material, section S4).

Final confirmation of successful COP attachment to the carbon surface was verified using transmission electron microscopy (TEM). Imaging with TEM revealed large dark activated carbon particles, out of focus in the foreground, with a shell of interwoven porous chains of COP material bound to the outer edges, in focus in the background (Figure 3). Confirming expected polymer formation and construction, the morphology of the COP used in this study and observed surrounding the GAC closely resembled that of the COP used in a previous study immobilizing nanoscale zero-valent iron (nZVI) into the pores of the COP (see Supplementary material, section S5) [21]. In addition, when TEM imaging was performed on particles just before COP attachment (AC-Mel) with only melamine bound to the surface, there was a noticeable difference in the material at the edges of the activated carbon. This material of melamine monomers attached appeared as an amorphous and nonporous coating, rather than a shell of a porous polymer matrix (see Supplementary material, section S5).

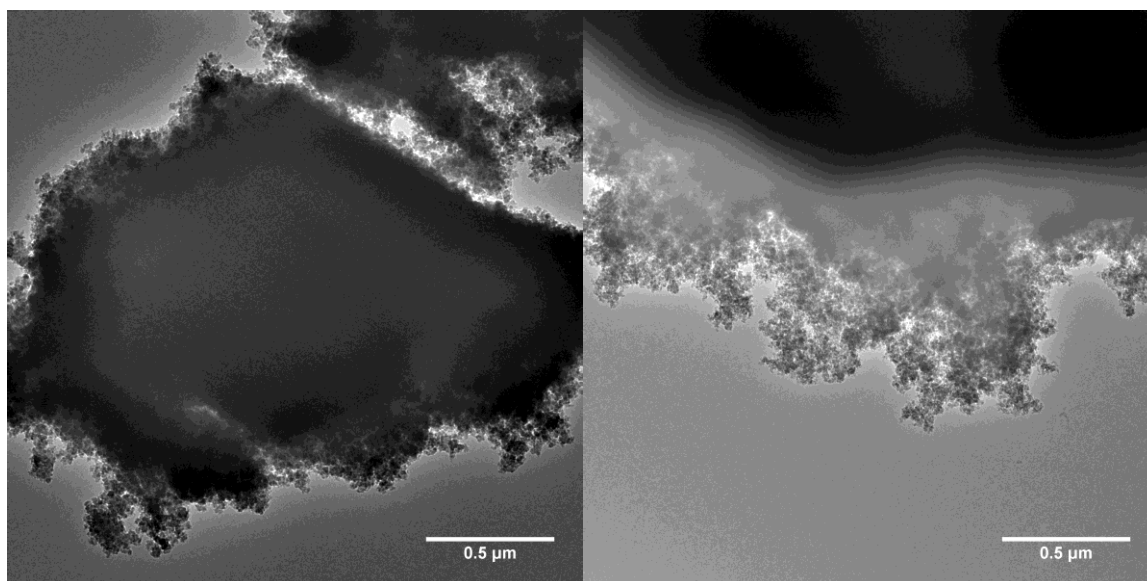


Fig. 3. TEM images of activated carbon particles with an attached shell of COP material.

Porosity and textural characterization was performed on the materials, using nitrogen adsorption-desorption isotherms at 77K (Figure 4). Activated carbon shows a typical type I isotherm, due to its microporous nature, with a very narrow pore size distribution centered around 2 nm (Figure 4A, black line). For comparison, the bare COP possesses a type II isotherm with a significant amount of macropores and a broad range of pore sizes (from 2.6 nm to 100 nm) (see Supplementary material, section S7). Not uncommon after oxidation of activated carbon [29,30], the porosity of AC-Ox is significantly reduced, reflecting the high content of oxidized functionalities blocking the pores (Figure 4B, red line). Melamine attachment and successive COP growing leads to an increase in surface area of 75 m²/g and 338 m²/g for AC-Mel and AC-COP, respectively. The increased porosity can be explained through partially re-opening of the blocked pores after oxidation (small pores centered around 2.3 nm), as well as newly generated larger pores from the COP structure itself (2.6 nm to 10 nm) but with a broader pore size distribution, reflecting the hybrid nature of the AC-COP.

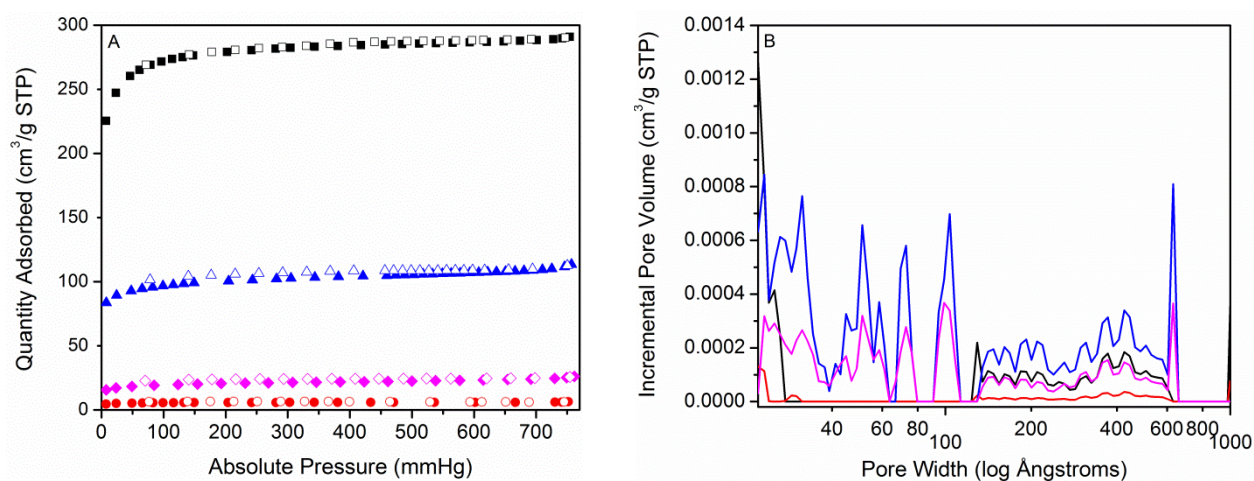


Fig. 4. A) N₂ adsorption (filled)-desorption (empty) isotherms at 77 K; B) Pore size/volume distribution by 2D-NLDFT (slit pores, N₂ carbon model) of AC (black), AC-Ox (red), AC-Mel (magenta), and AC-COP (blue).

Finally, in order to test the preliminary efficacy of the produced material for environmental clean-up and contaminant removal/adsorption, in separate batch tests, AC-COP was mixed with the azo dye naphthol blue

black and cadmium. Iron was also tested, to assess possibility for future application of impregnating nZVI into AC-COP for subsequent contaminant degradation. For comparison, AC and COP-19 were also examined. After only 30 minutes, AC-COP and COP-19 exhibited a high affinity for adsorbing the azo dye, very similar to previously reported results using COP-19 [21]; while the AC did not, each adsorbing up to 23.6, 23.8, and 5.6 mg/g, respectively (Figure 5A). The high adsorption of an azo dye to the COP material has been exhibited before [21], and can best be explained by the phenomenon known as pi-pi stacking, which has been observed before between aromatic compounds [31–33]. After 72 hours, adsorption for cadmium was 5.6, 4.5, and 5.3 mg/g for AC-COP, COP-19, and AC, respectively (Figure 5B); similar results have been observed adsorbing cadmium with carbon nanotubes [34]. Interestingly, COP-19 exhibited the lowest cadmium adsorption, however, there was a synergistic effect observed between the AC and COP-19 in AC-COP, which adsorbed the highest amount of cadmium. After 72 hours, adsorption for iron was 8.7, 12.4, and 11.5 mg/g for AC-COP, COP-19, and AC, respectively (Figure 5C). Although, having lower adsorption numbers, AC-COP continued to increase adsorption of iron over a 3 day period, while AC and COP-19 saw a release of iron back into solution after 8 hours, indicating that AC-COP is permanently trapping iron, making it an ideal candidate of nZVI I impregnation for future studies on contaminant degradation. Ultimately, the rate constants of adsorption were determined using pseudo first-order kinetics [35]. The rate constants for azo dye adsorption were 0.0678, 0.1304, and 0.2211 h⁻¹ for AC, COP-19, and AC-COP, respectively; the rate constants for cadmium adsorption were 0.0012, 0.0016, and 0.0011 h⁻¹ for AC, COP-19, and AC-COP, respectively; and the rate constants for iron adsorption were 0.0203, 0.0124, and 0.0121 h⁻¹ for AC, COP-19, and AC-COP, respectively. Preliminary adsorption batch tests indicate the ability of the developed material to be used effectively in environmental applications.

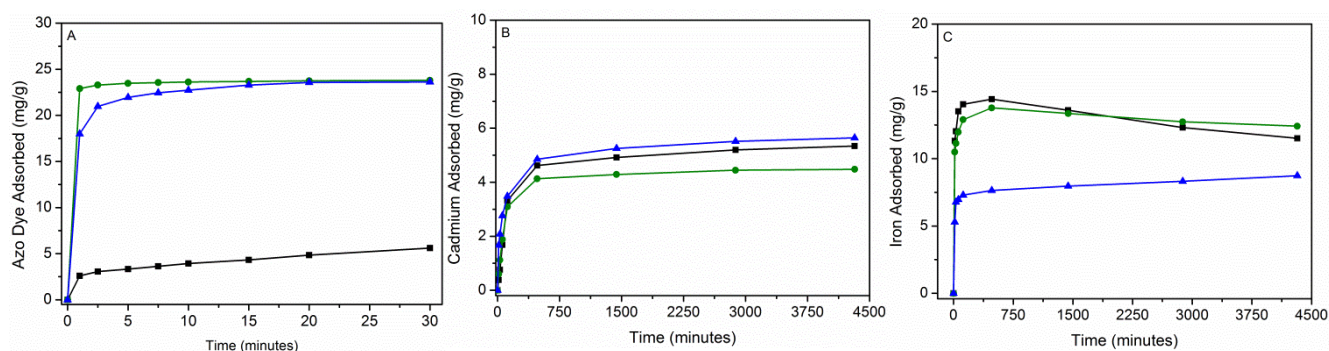


Fig. 5. Adsorption isotherms of A) azo dye naphthol blue black; B) cadmium; and C) iron adsorbed by AC-COP (blue), COP-19 (green) and AC (black). Material dosage: 0.5 g/L; initial pH: 6.5; C_0 -azo dye: 40 μ M; C_0 -Cd: 10 mg/L; C_0 -Fe: 10 mg/L.

4. Conclusions

In summary, although various types of activated carbons were tested, it was determined that very pristine or crude carbons are not adequate for surface modification. If the carbon is too pristine, then the surface oxidation becomes extremely difficult; and, if the carbon is too crude, then the structural integrity of the carbon becomes a concern. Therefore, a carbon with a moderate amount of defects that possesses a high surface area is an optimal material for polymer functionalization of GAC. Defects initially present in the carbon leads to ease of oxidation of the surface, creating functional carboxyl groups for further modification. These carboxyl groups are easily converted in to acyl chlorides via thionyl chloride treatment. From the acyl chloride, monomer attachment, in the form of melamine, is then possible. Monomers such as this can provide the building infrastructure for the construction of a polymeric network around the surface of the carbon. In this manner, we successfully created a covalent organic polymer shell around activated carbon granules, the first time this type of porous framework has been attached to the surface of activated carbon. From here, we expect to establish a “library” of surface modified polymer attached carbons for various environmental applications; ranging from heavy metal adsorption,

to water contamination treatment, to CO₂ adsorption, to biogas purification. Consequently, following the successful initial results of contaminant adsorption by AC-COP, the environmental applications are promising.

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